

Supporting Information

Life Cycle Assessment and Techno-economic Analysis of Thermochemical Conversion Technologies Applied to Poultry Litter with Energy and Nutrient Recovery

Raaj R. Bora,^{1,†} Musuizi Lei,^{1,†} Jefferson W. Tester,^{1,3} Johannes Lehmann,^{2,3} Fengqi You^{1,3,*}

¹ *Robert Frederick Smith School of Chemical and Biomolecular Engineering, Cornell University, Ithaca, New York 14853, USA*

² *Soil and Crop Sciences, School of Integrative Plant Science, College of Agriculture and Life Sciences, Cornell University, Ithaca, New York 14853, USA*

³ *Atkinson Center for a Sustainable Future, Cornell University, Ithaca, New York 14853, USA*

This file contains: 27 pages, 5 sections, 9 tables and 5 figures.

* Corresponding author. E-mail: fengqi.you@cornell.edu

† Raaj R. Bora and Musuizi Lei contributed equally to this work

List of Sections

- S1 - Description of thermochemical technologies
- S2 - Mass and energy balances
- S3 - Life cycle inventory (LCI) data and sensitivity analysis parameters
- S4 - Life cycle impact assessment (LCIA) categories and factors
- S5 - Techno-economic analysis parameters

List of Tables

- Table S1 - Major components of each system considered in the analysis.
- Table S2 - Product distribution and element fate for bioconversion technologies.
- Table S3 - Life cycle inventory associated with 1000 kg of wet poultry litter treated.
- Table S4 - Comparison between soil products.
- Table S5 - Selected parameters along with their base values and ranges used for the sensitivity analysis.
- Table S6 - Characterization Factors (CF) used for the 15 mid-point categories.
- Table S7 - Normalization factors used for the damage categories.
- Table S8 - Life cycle impact assessment (LCIA) of thermochemical conversion of poultry litter
- Table S9 - Additional sensitivity analysis based on capital costs.

List of Figures

- Figure S1 - Energy and N flow for slow pyrolysis and HTL processes for 1000 kg wet poultry litter as a representation of the various balances used in this analysis.
- Figure S2 - Mass and energy balance and C and N flows for 1 kg wet biomass via HTL process.
- Figure S3. Calculated LCIA midpoint contribution for the different technologies
- Figure S4 - Calculated LCIA normalized scores for the different technologies.
- Figure S5 - Sensitivity analysis of the LCA and TEA results for the three technologies - slow pyrolysis (SP), hydrothermal liquefaction (HTL) and gasification (GA).

S1. Description of thermochemical technologies

Table S1. Major components of each system considered in the analysis. The CHP is considered to have pollutant removal technology installed, with NO_x removal efficiency at the same level as that of the SCR (0.9). The gas phase for the hydrothermal technologies consists of the gases produced by the technologies themselves and from the AD used to treat their aqueous phases.

	Rearing of poultry	Poultry litter collection	Drying	Main products	Gas phase	Oil phase	Aqueous phase	Solid phase
Direct land application system	No	Tractor & trailer	No	N/A	N/A	N/A	N/A	Direct soil application of poultry litter
Slow pyrolysis system	No	Tractor & trailer	Yes	Gas, Bio-oil, Biochar	Combusted on site with SCR	Upgraded to fuels	N/A	Soil application of biochar
Fast pyrolysis system	No	Tractor & trailer	Yes	Gas, Bio-oil, Biochar	Combusted on site with SCR	Upgraded to fuels	N/A	Soil application of biochar
Gasification system	No	Tractor & trailer	Yes	Gas, Biochar	CHP	N/A	N/A	Soil application of biochar
HTC system	No	Tractor & trailer	No	Gas, Aqueous, Hydrochar	CHP	N/A	Digested (AD)	Soil application of hydrochar and digestate
HTL system	No	Tractor & trailer	No	Gas, Aqueous, Bio-crude, Hydrochar	CHP	Upgraded to fuels	Digested (AD)	Soil application of hydrochar and digestate
SCWG system	No	Tractor & trailer	No	Gas, Aqueous, Syngas	CHP	N/A	Digested (AD)	Soil application of digestate

S1.1. Pyrolysis of poultry litter

Pyrolysis is one among the thermochemical technologies that are considered in this analysis (Table S1). Pyrolysis is the process involving thermal degradation of organic matter at elevated temperatures in an inert atmosphere. Pyrolysis processes can be split into slow pyrolysis and fast

pyrolysis depending on reaction temperature and duration¹. Two typical studies are selected for slow pyrolysis and fast pyrolysis respectively as the basis to build mass and energy flows.²⁻⁴ In both studies, the elements' fate and distribution among product fractions are provided in detail (as summarized in Table S2). In the 400-degree Celsius slow pyrolysis, 1 kg fresh feedstock results in 0.44 kg biochar, which contains 51% of feedstock energy, 63% carbon and 59% nitrogen. In the fast pyrolysis studies, nearly half of the primary energy, carbon and nitrogen go into bio-crude.

The process scheme for slow and fast pyrolysis is almost identical in our study. First, after collection, pre-treatment and drying, the biomass is sent to the reactor. The biochar produced is directly sent for soil application and the bio-oil is sent to a refinery plant to produce diesel and gasoline products. The syngas produced is burnt onsite in a boiler to supply energy for internal consumption. The low-energy content condensable vapor (aqueous phase), as well as the filtered water from the drying processes are regarded as wastes and are sent for wastewater treatment. The detailed emissions associated with each unit process are provided in Table S3.

S1.2. Gasification of poultry litter

Gasification is carried out at elevated temperatures and in a restricted oxygen environment. The fed biomass is converted into a gaseous energy carrier called “syngas” or “product gas”. In the past, poultry litter has been regarded as an unconventional fuel to gasify, because the high ash content in poultry litter requires low reaction temperatures to avoid ash agglomeration, but low temperatures only lead to low gas yields. However, a recent study has successfully achieved an energy efficiency of up to 89% by blending poultry litter with limestone at 8% and setting the temperature to 800 degree Celsius. This great energy ratio is aided by calcite addition, which reduces the bed agglomeration risk and permits high operation temperatures. It deserves mention that this experiment is performed at a load rate of less than 1 kg/h and the large-scale production stability remains unknown.⁵ This limestone does not serve as a CO₂ absorbent but as a means to improve the product yield. For gasification, all the available carbon in the feedstock would eventually be converted to CO₂. The emissions could be reduced either with some form of carbon capture, or as in the case of the pyrolysis technologies, through biochar which sequesters a part of the carbon.⁶ Considering there are other poultry litter gasification studies which have successfully simulated the process with a load rate of 1500 kg/h, we assume the large-scale production is

feasible.⁷ In this LCA, the produced syngas from gasification (with air as the gasifying agent) is sent to an existing CHP by using a biogas truck, based on a typical transportation method in distributed biogas systems.^{8,9} For simplification, we assume that there are no stray gas emissions during the gasification process itself and the syngas transportation. The pre-drying water and gasification ash are regarded as industrial waste. The biochar produced from the gasification process is assigned a similar economic value as the pyrolysis biochars (and this may be conservative as the concentration of nutrients is comparatively higher than the other biochars). As with pyrolysis biochars, gasification biochar cannot be considered as a nitrogen fertilizer substitute owing to the low levels of plant-available nitrogen present in the biochar. We assume that the biochar is the source of some organic carbon to the soil as with pyrolysis biochars, though the quantity of biochar produced through gasification is much lower.¹⁰

S1.3. Hydrothermal methods for treatment of poultry litter

Hydrothermal processes are novel technologies to convert wet biomass into biofuels using sub-critical water as the solvent. Unlike in the “dry” thermal conversion technologies, water provides an excellent environment, and acts as a reactant and solvent for the typical reactions associated with the hydrothermal technologies. To ensure consistency, a comparative study for the three hydrothermal technologies on poultry litter is utilized and cited. Water is added to the poultry litter for diluting the feed. Since poultry litter is a relatively dry organic feedstock, this dilution is currently required to make it suitable for the hydrothermal technologies, which function best with wetter feeds. An alternative would be to consider recycled water streams for the dilution purposes if possible. It is interesting to note that over half of the mass, 60% of the Nitrogen (N), and 90% of the Potassium (K) go into the aqueous phase. In our process scheme design, all the aqueous primary products from hydrothermal processes are further sent to AD-CHP units to recover energy and nutrients. The hydrochar is sent for direct soil application. Bio-crude from HTC and SCWG are regarded as waste, and only HTL bio-crude goes through hydrotreatment to produce transportation fuels because of the comparatively larger volumes produced. Similarly, only the SCWG gas is sent to CHP, and the HTL and HTC gases are burnt on site for internal utilization of heat produced. This comes with its set of emissions too.¹¹

S2. Mass and energy balances

The mass, energy, carbon, and nitrogen output distribution for the six technologies are represented through the following tables and figures (Table S2, Figure S1-S2). The feed for all of these processes is considered to have the same properties (as described in the main text).

Table S2. Product distribution and element fate for bioconversion technologies. The summed ratios do not necessarily add up to 100% because of measurement errors or heat losses.

Studies		ref ²	ref ^{3, 4, 12}	ref ⁵	ref ¹¹	ref ¹¹	ref ¹¹
Technology		Slow pyrolysis	Fast pyrolysis	Gasification	HTC	HTL	SCWG
Mass Distribution	solids	0.55	0.42	0.20	0.35	0.17	0.12
	bio oil/crude	0.15	0.27	-	0.03	0.17	0.02
	aqueous phase	0.10	0.18	-	0.50	0.54	0.55
	syngas	0.20	0.13	0.80	0.12	0.12	0.31
Energy Distribution	solids	0.48	0.32	-	0.70	0.31	0.18
	bio oil/crude	0.29	0.49	-	0.06	0.36	0.04
	aqueous phase	0.04	-	-	0.14	0.10	0.05
	syngas	0.06	0.03	0.70	0.05	0.05	0.50
Carbon Distribution	solids	0.63	0.30	0.11	0.27	0.28	0.18
	bio oil/crude	0.25	0.48	-	0.09	0.36	0.05
	aqueous phase	0.05	0.17	-	0.53	0.25	0.30
	syngas	0.07	0.05	0.89	0.11	0.11	0.47
Nitrogen Distribution	solids	0.59	0.24	0.25	0.30	0.12	0.07
	bio oil/crude	0.07	0.52	-	0.05	0.17	0.05
	aqueous phase	0.20	0.15	-	0.60	0.58	0.63
	syngas	0.16	0.10	0.75	0.02	0.02	0.20

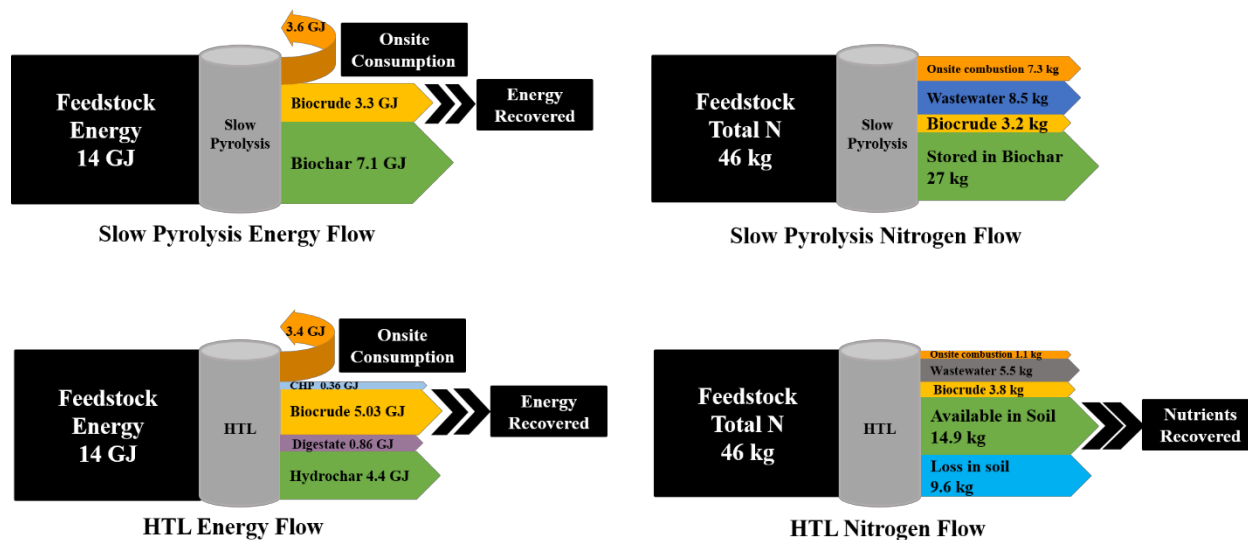


Figure S1. Energy and N flow for slow pyrolysis and HTL processes for 1000 kg wet poultry litter as a representation of the various balances used in this analysis.

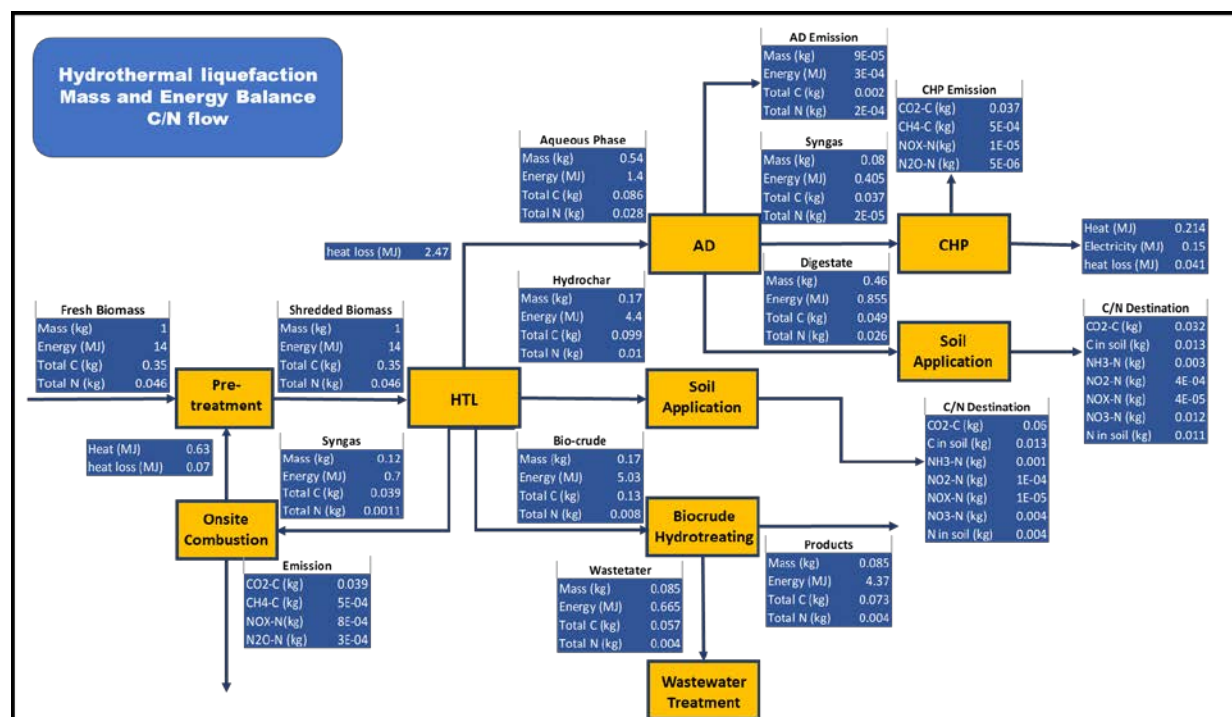


Figure S2. Mass and energy Balance and C/N flow for 1 kg wet biomass via HTL process. The external energy and material as well as emissions from other elements are not shown in this figure.

S3. Life cycle inventory (LCI) data and sensitivity analysis parameters

S3.1. Introduction

This section covers the LCI data and describes some of its sources as well as a brief description of the processes themselves. It also includes a list of parameters and assumptions used in the sensitivity analysis. Table S3 contains the LCI data used for calculating the LCIA scores. Each row in this table is derived from one or more processes, which have been grouped together, based on what they consume or produce.

Table S3. Life cycle inventory associated with 1000 kg of wet poultry litter treated

Item	Unit	Slow Pyrolysis	Fast Pyrolysis	Gasification	HTC	HTL	SCWG	Reference Case
Inputs from Technosphere								
trailer	t.km	5.00E+00	5.00E+00	5.00E+00	5.00E+00	5.00E+00	5.00E+00	5.00E+00
truck	t.km	7.80E+01	7.76E+01	0.00E+00	1.18E+02	1.21E+02	1.11E+02	0.00E+00
gas truck	t.km	0.00E+00	0.00E+00	6.43E+00	0.00E+00	0.00E+00	3.10E+00	0.00E+00
electricity	kWh	2.76E+02	2.97E+02	-7.50E+02	1.98E+02	2.49E+02	-3.19E+02	0.00E+00
heat	MJ	4.18E+02	9.08E+02	-3.49E+03	-8.05E+02	-1.07E+02	-1.31E+03	0.00E+00
natural gas	m ³	1.36E+01	2.50E+01	0.00E+00	0.00E+00	1.93E+01	0.00E+00	0.00E+00
tap water	kg	4.46E+01	8.18E+01	0.00E+00	0.00E+00	6.32E+01	0.00E+00	0.00E+00
limestone	kg	0.00E+00	0.00E+00	8.70E+01	0.00E+00	0.00E+00	0.00E+00	0.00E+00
lubricating oil	kg	0.00E+00	0.00E+00	2.90E-01	1.00E-02	1.00E-02	1.60E-01	0.00E+00
steel	kg	8.00E-01	8.00E-01	8.00E-01	8.00E-01	8.00E-01	8.00E-01	0.00E+00
iron	kg	1.00E-02	1.00E-02	1.00E-02	1.00E-02	1.00E-02	1.00E-02	0.00E+00
aluminum	kg	1.00E-02	1.00E-02	1.00E-02	1.00E-02	1.00E-02	1.00E-02	0.00E+00
concrete	kg	2.51E+00	2.51E+00	2.51E+00	2.51E+00	2.51E+00	2.51E+00	0.00E+00
AD	plant eq	0.00E+00	0.00E+00	0.00E+00	4.65E-06	3.33E-06	1.66E-06	0.00E+00
CHP	plant eq	0.00E+00	0.00E+00	1.15E-05	5.52E-07	3.94E-07	7.02E-06	0.00E+00
boiler	plant eq	1.62E-06	8.22E-07	0.00E+00	3.92E-06	1.73E-06	1.45E-06	0.00E+00
refinery plant	plant eq	4.92E-09	8.86E-09	0.00E+00	0.00E+00	5.58E-09	0.00E+00	0.00E+00
slurry spreading	m ³	4.60E-01	3.50E-01	0.00E+00	7.00E-01	6.40E-01	5.80E-01	1.67E+00

Direct Emission								
NM VOC	kg	2.15E-03	1.54E-03	8.00E-02	1.00E-02	6.69E-09	4.00E-02	0.00E+00
NH ₃	kg	0.00E+00	0.00E+00	0.00E+00	3.19E+00	3.56E+00	3.96E+00	1.20E+01
CO ₂	kg	6.40E+02	9.80E+02	1.17E+03	1.28E+03	1.28E+03	1.28E+03	1.20E+03
CH ₄	kg	2.00E-02	2.00E-02	9.00E-01	9.10E-01	4.20E-01	9.60E-01	3.01E+01
NO _x	kg	1.79E+00	1.12E+00	1.30E-01	8.80E-01	3.20E-01	1.10E-01	1.90E-01
N ₂ O	kg	3.00E-01	1.90E-01	2.00E-01	8.00E-01	7.80E-01	8.80E-01	8.90E-01
CO	kg	5.00E-02	4.00E-02	1.88E+00	3.20E-01	1.60E-01	1.08E+00	0.00E+00
P to river	kg	0.00E+00	0.00E+00	0.00E+00	9.00E-01	3.21E+00	8.20E-01	1.36E+00
N to ground	kg	0.00E+00	0.00E+00	0.00E+00	8.30E+00	6.50E+00	6.45E+00	3.57E+01
Outputs from Technosphere: Avoided Products								
gasoline	kg	-2.56E+01	-4.69E+01	0.00E+00	0.00E+00	-3.63E+01	0.00E+00	0.00E+00
diesel	kg	-3.44E+01	-6.31E+01	0.00E+00	0.00E+00	-4.87E+01	0.00E+00	0.00E+00
fertilizer N	kg	-5.16E+00	-3.41E+00	-2.19E+00	-3.11E+01	-2.42E+01	-2.42E+01	-1.61E+01
fertilizer K ₂ O	kg	-2.41E+01	-2.33E+01	-1.98E+01	-5.78E+00	-5.62E+00	-4.15E+00	-4.98E+00
fertilizer P ₂ O ₅	kg	-2.71E+01	-2.74E+01	-2.33E+01	-3.94E+00	-1.41E+01	-3.59E+00	-5.91E+00
electricity	kWh	0.00E+00	0.00E+00	-1.21E+03	-5.83E+01	-4.17E+01	-6.85E+02	0.00E+00
heat	MJ	4.62E+02	-2.12E+02	-6.27E+03	-1.25E+03	-6.34E+02	-4.17E+03	0.00E+00
Waste Treatment								
ash	kg	0.00E+00	0.00E+00	1.57E+02	0.00E+00	0.00E+00	0.00E+00	0.00E+00
waste mineral oil	kg	0.00E+00	0.00E+00	2.90E-01	1.00E-02	1.00E-02	1.60E-01	0.00E+00
wastewater	kg	3.40E+02	4.54E+02	2.00E+02	0.00E+00	8.50E+01	0.00E+00	0.00E+00

S3.2. Soil Products

In the direct land application case, only collection and slurry spreading activities are involved. The environmental burdens mostly come from GHG emission and leaching.¹³⁻¹⁹ It has been estimated that over 30% of N applied on land will be converted into ammonia,²⁰ and 35% will leach into groundwater. Additionally, land applied with organic fertilizer also tends to emit 32.2% more N₂O

than synthetic N fertilizer.¹⁵ 21% of the Phosphorus (P) contained will leach into surface water. The remaining P is precipitated in the soil.^{21, 22}

In this study, it is assumed that 90% of biochar carbon will be stably stored, all the biochar P and K will finally be taken up by plants and per ton, biochar will help mitigate 0.3933 kg N₂O emissions. The assumptions are consistent with other field studies conducted on biochar.²³⁻²⁵ For the fertility improvement effect, instead of assuming short-term effects, here we assume biochar will keep this function for three years as a base scenario, which could finally save 14.6 kg N, 8.9 kg K₂O and 6.2 kg P₂O₅ fertilizers.²⁴⁻²⁶

In contrast, hydrochar and digestate receive far less attention as soil amendments principally due to their poor performance in carbon sequestration²⁷⁻²⁹. 68-88% and 48-77% hydrochar carbon were reported to be lost after one-year laboratory incubation and one-year field incubation study, respectively. There is no strong evidence that hydrochar will improve fertility except from the nutrients themselves.³⁰ Table S4 summarizes the properties of the different soil products considered in a comparative form.

Table S4. Comparison between soil products^{23, 25, 27, 29}

	Biochar	Hydrochar	Digestate	Fresh Poultry Litter
Production process	High temperature "dry" pyrolysis and gasification	"Wet" pyrolysis, energy saving	From AD process or co-digestion	Directly from farms
Available of nutrients	N is unavailable, other nutrients are available	Most of nutrients are available, but certain parts cannot be utilized.	Most of nutrients are available, but certain parts cannot be utilized.	Most of nutrients are available, but certain parts cannot be utilized.
Emissions	CO ₂ is stably stored, soil N ₂ O and NH ₃ emissions are mitigated. No CH ₄ emissions observed.	Risk of stimulating microbial activity and increase in N ₂ O emissions. Also C-based emissions.	N is highly unstable, with high NH ₃ and N ₂ O emissions.	N is highly unstable, with high NH ₃ and N ₂ O emissions. High CH ₄ emissions observed.
Leaching	Slowly degradable, but leaching is not considered.	Excess P and N leaching to surface water and underground water.	Excess P and N leaching to surface water and underground water.	Excess P and N leaching to surface water and underground water.

S3.3. Energy production processes

The pre-treatment heat required for drying and the heat required to reach the reaction temperature is calculated based on the work by Wang et al.³¹ The operation electricity demand for all the processes is assumed as 0.24 kwh per kg wet biomass.^{32, 33} The onsite combustion of syngas and biocrude is assumed at an energy efficiency of 60%. Here we assume selective catalytic reduction (SCR) is present in the boiler, which reduces 90% N emissions. The ratio of C and N compounds emissions are mainly extracted from the Ecoinvent boiler emission data. Similarly, the CHP and AD data, including the energy consumption and emissions, are all from Ecoinvent.³⁴ The data concerning bio-oil upgrading and hydrotreating are mainly cited from a fuel upgrading study and a comparative study, where similar final product yields are achieved (but their components vary).³³ Although the system boundary ends with the diesel and gasoline ready to be sold, the carbon dioxide emissions from bio-fuel burning are still included from the perspective of maintaining carbon balance.

S3.4. Transportation, electricity and construction

The transportation distances are assumed as 50 km for all the facility-to-farm processes as well as the distances between the different facilities themselves. This value is higher than other studies and one reason is that poultry litter is produced in small volumes and the conversion plant is assumed as relatively centered. The distance estimated here is also suggested by poultry fertilizer spatial allocation analysis.³⁵ The systems end with the diesel and gasoline ready to be sold, and 50 km is taken as the value for plant to market distance.

The electricity produced from CHP is assumed to be consumed onsite, so no electricity transmission is involved. Data related to waste collection is hard to evaluate, and here we assume a small tractor and trailer for the collection of poultry litter on the farm. Every collection activity is estimated to last for 30 minutes based on the dairy manure collection system for reference.³⁶ The materials for construction of all the plants are assumed the same and the required materials for a gasification plant with a capacity of 10 ton/h are used as the reference case. Construction materials in different cases are then scaled according to capacity.

S3.5. Sensitivity Analysis Parameters

The following table specifies the selection of the values of various parameters for the sensitivity analysis and our assumptions involved, as not all the literature that is cited was exclusively based on poultry litter.

Table S5. Selected parameters along with their base values and ranges used for the sensitivity analysis. References for the values and ranges are provided in the parameter column. (In the units' column, feed is referred to on a dry basis).

Parameter	Technology	Base value	Range	Units
Electricity consumption	Slow pyrolysis	275	100–500	kWh
Biochar efficacy period	Slow pyrolysis	3	1–9	years
Biochar yield ³⁷	Slow pyrolysis	0.55	0.4–0.6	kg/kg feed
Bio-oil yield	Slow pyrolysis	0.12	0.06–0.2	kg/kg feed
Transportation distance ³⁴	Slow pyrolysis	78	0–150	km
Hydrochar yield	HTL	0.17	0.1–0.3	kg/kg feed
N-uptake amount ²⁵	HTL	0.75	0.25–1	kg/kg-N feed
Biocrude yield ³³	HTL	0.17	0.1–0.3	kg/kg feed
Transportation distance ³⁴	HTL	117.5	20–200	km
CHP energy efficiency ³⁴	Gasification	0.9	0.6–0.95	kg/kg feed
Syngas yield ²⁹	Gasification	0.8	0.7–0.9	kg/kg feed
Biomass transportation ³⁴	Gasification	10	5–100	km
Biochar value ^{26, 38, 39}	Slow pyrolysis	182	0–300	\$/1000 kg
CO ₂ price ⁶	Slow pyrolysis	20	0–500	\$/1000 kg CO ₂ -eq
Biogas transport ³⁴	Gasification	10	0–20	km

S4. Life cycle impact assessment (LCIA) categories and factors

S4.1. LCIA parameters

The IMPACT 2002+ framework consists of fifteen mid-point categories and four damage-oriented or end-point categories. The four damage-oriented impact categories considered (along with their units) are: human health (DALY), ecosystem quality (PDF*m²*year), climate change (kg CO₂-eq) and resource depletion (MJ primary).⁴⁰ DALY (Disability-Adjusted Life Years) characterizes the disease severity, accounting for both mortality (years of life lost due to premature death) and morbidity (the time of life with lower quality due to an illness). PDF*m²*y (Potentially Disappeared Fraction of species over a certain amount of m² during a certain amount of years) is the unit to measure the impacts on ecosystems. GWP (Global Warming Potential) of 100 years is measured in units of kg CO₂-eq for the climate change category. MJ (Mega Joules) measures the amount of energy extracted or needed to extract the resource.⁴¹

For human health, in DALY (disability adjusted life year), the mid-points are carcinogens, non-carcinogens, respiration inorganics, ionizing radiation, ozone layer depletion, respiratory organics. For ecosystem quality, which is presented as PDF*m²*year, the mid-points include aquatic ecotoxicity, terrestrial ecotoxicity, terrestrial acid/nutri, aquatic acidification and aquatic eutrophication. The climate change category (in kg CO₂ eq) only consists of global warming, and the resources category (in MJ primary) includes non-renewable energy and mineral extraction.⁴⁰ The weights and normalization factors used for these categories can be found in Table S6-S7. In this study, the LCI data is initially utilized to calculate the values for each of the fifteen midpoint categories.⁴⁰ Then, the midpoint values are converted into the four damage categories by multiplying with the respective characterization factors. Ultimately, the endpoint category values are normalized into points by dividing them by the associated normalization factors.

Table S6. Characterization Factors (CF) used for the 15 mid-point categories.⁴¹

Category	Factor	Units	Category	Factor	Units
Carcinogens	2.80E-06	DALY/kg C2H3Cl eq	Terrestrial acidification/nutrication	1.04E+00	PDF•m ² •yr /kg SO ₂ eq
Non-carcinogens	2.80E-06	DALY/kg C2H3Cl eq	Land occupation	1.09E+00	PDF•m ² •yr /m ² org.arable land-yr
Respiratory inorganics	7.00E10-4	DALY/kg PM2.5 eq	Aquatic acidification	8.82E-03	PDF•m ² •yr /kg SO ₂ eq

Ionizing radiation	2.10E-10	DALY/Bq C-14 eq	Aquatic eutrophication	1.14E+01	PDF•m ² •yr /kg PO4 P-lim
Ozone layer depletion	1.05E-03	DALY/kg CFC-11 eq	Global warming	1.00E+00	kg CO ₂ -eq/kg
Respiratory organics	2.13E-06	DALY/kg C ₂ H ₄ eq	Non-renewable energy	1.00E+00	MJ primary/kg
Aquatic ecotoxicity	5.02E-05	PDF•m ² •yr /kg TEG water	Mineral extraction	1.00E+00	MJ surplus/kg
Terrestrial ecotoxicity	7.91E-03	PDF•m ² •yr /kg TEG soil			

Table S7. Normalization factors used for the damage categories.

Human Health (DALY/pt)	Ecosystem Quality (PDF•m ² •yr/pt)	Climate Change (Kg CO ₂ -eq/pt)	Resource Depletion (MJ/pt)
0.0071	13,700	9,900	152,000

S5. Techno-economic analysis parameters

S5.1. Capital costs

The methods to estimate the pyrolysis and gasification capital costs are derived from the studies by Bridgewater et al, where the capital costs and plant size relationships are separately provided for fast pyrolysis plants and gasification plants. These equations have considered the learning effects, which refers to the phenomenon that the cost reduces with more units built and more experience accumulated. In this study, a learning factor of 50% has been assumed, corresponding to 10 installations of a novel process. The plant costs are also updated according to CEPCI from 394.1 in 2000 to 567.5 in 2017. The pyrolysis (including slow and fast pyrolysis) and gasification plant costs are calculated by the following equations:

$$TPC_{pyrolysis} = 39.7 \times (Q \times 1000)^{0.6194} \quad (S1)$$

$$TPC_{gasification} = 92 \times (Q \times 1000)^{0.6384} \quad (S2)$$

Where, TPC is the total plant cost ($kUSD_{2018}$) and Q is the feedstock input rate (in dry ton/hour). In this study, the average plant size is set to be 30 ton/day, which was calculated by equally

assigning the total amount of poultry litter in NYS into 10 plants. The capital cost calculated for pyrolysis and gasification plants are \$3.2MM and \$8.73MM, respectively. Besides, since the above equations do not include pretreatment modules, and the capital cost of pre-treatment part is around 27% of the main reactor capital cost according to the same study, the total capital costs for pyrolysis and gasification become \$4.17MM and \$11.1MM, respectively. The equivalent capital cost is calculated based upon a 20-year lifetime and a discount rate of 10%.

$$Annual_capital = \frac{Total_capital}{\frac{1}{r} - \frac{1}{r(1+r)^n}} \quad (S3)$$

The annual capital costs are \$0.49MM and \$1.31MM, and the costs for unit ton feedstock are \$44 and \$119 for pyrolysis and gasification, respectively. The capital cost calculations for the hydrothermal technologies follow the methods adopted by Van Doren et al.⁴²

S5.2. O&M costs

The O&M costs include fixed operating costs and variable operating costs. The fixed part mainly consists of operation labor cost (a function of plant size), maintenance labor cost (1% the total capital cost), overheads (2%), maintenance materials (3%), taxes and insurance (2%) and other fixed costs (1%). The variable part primarily includes, in this case, the electricity and natural gas cost as well as waste handling. For gasification, there is limestone involved as catalysts. The wastewater cost is \$0.73/1000 kg, electricity price is \$0.06/kwh, natural gas is \$0.022/MJ, limestone price is \$30/1000 kg, and the labor rate for this size plant is 0.02 M\$/annum. The O&M costs for the six technologies are calculated to be \$62, \$102, \$155, \$86, \$59 and \$130 per 1000 kg wet poultry litter, respectively.⁴³⁻⁴⁵

S5.3. Secondary treatment

Secondary treatment refers to the processes converting primary products into final products including the AD process, CHP process, hydrotreating and oil upgrading. The hydrotreating plant and upgrading plants are considered as newly constructed plants in this study because the oil refinery plants are mostly built around oil wells and are not as abundant and widely distributed as AD and CHP. For AD and CHP, we consider utilization of existing facilities. However, in this study, we utilize existing ADs which are assumed to have spare capacity, and the government is

assumed to be crediting a gate-fee for the AD plants to encourage co-digestion^{42, 46}. Here, we assume the poultry litter plant stakeholders do not need to pay for AD digestate disposal. The cost for CHP is \$0.0077/MJ energy produced for a relatively large-sized unit.²⁷ It is interesting to observe that the cost for bio-oil upgrading is much higher than bio-crude hydrotreating according to a TEA study. The upgrading cost is \$10.3/GJ oil produced while the hydrotreating cost is \$2.4/GJ.³³ The low cost for bio-crude hydrotreating holds the key if hydrothermal technologies are to economically outweigh “dry” technologies in terms of the production costs. The costs for selective catalytic removal (SCR) implementation are also calculated.⁴⁷

S5.4. Transportation

The road transportation costs follow the equation:

$$T_c = 4.8 + 0.094 * D \quad (S4)$$

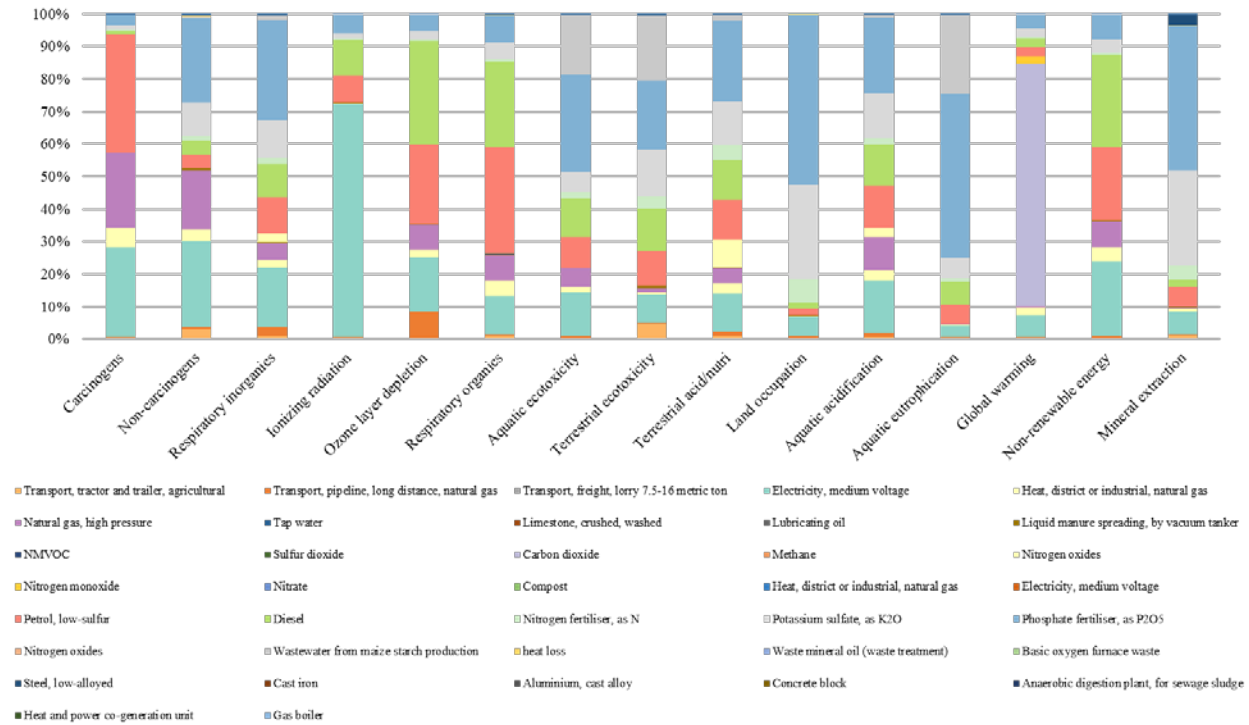
where 4.8 is the load and unload charge per ton, and 0.094 is the shipping cost per ton-km. All the value are in 2009 USD, and are converted to the corresponding 2017 values.²⁶ The biogas transportation cost is calculated to be \$0.07/Nm³ per km.⁴⁸

S5.5. Governmental credits

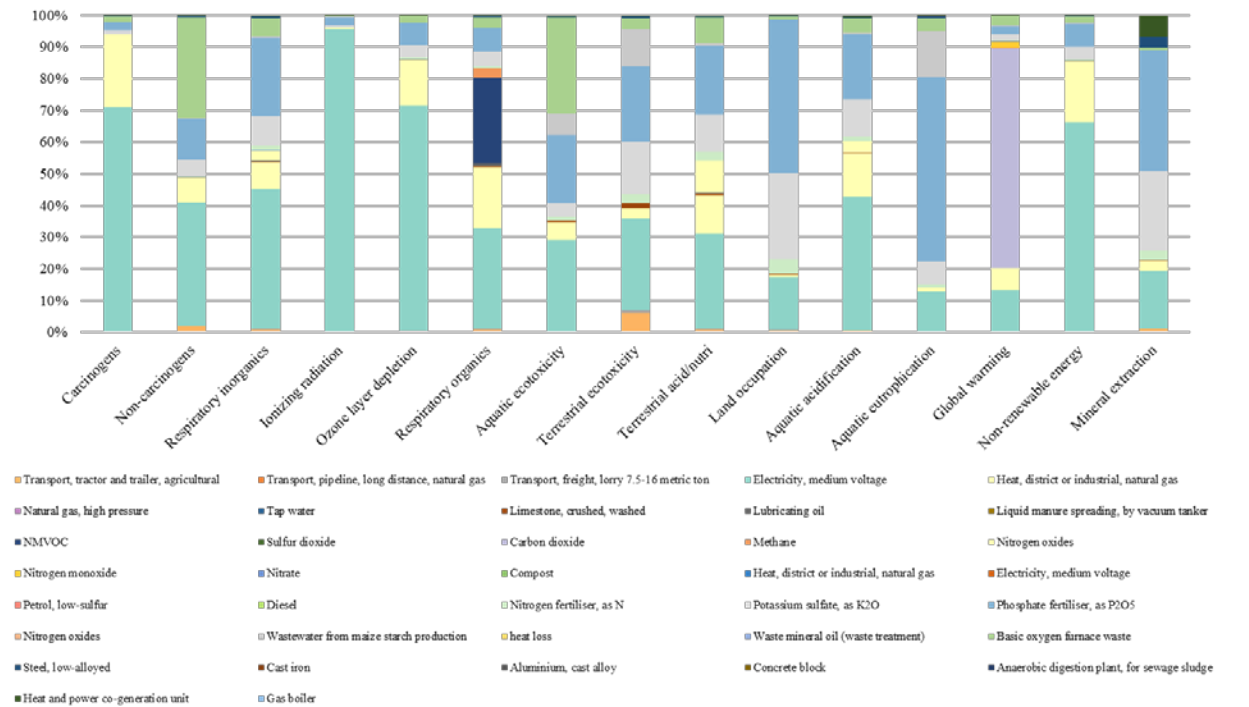
The governmental credits for carbon tax remain very low. In 2017 and 2018, the allowance was \$7/t CO₂e to \$16/t CO₂e, respectively in the European Union⁴⁹ and there were no credits in the USA. Here we estimate an average value of \$20/t CO₂e considering the predicted scenarios in the cited references. The avoided compost fees and the credits for CHP are not considered.

S6. Additional results

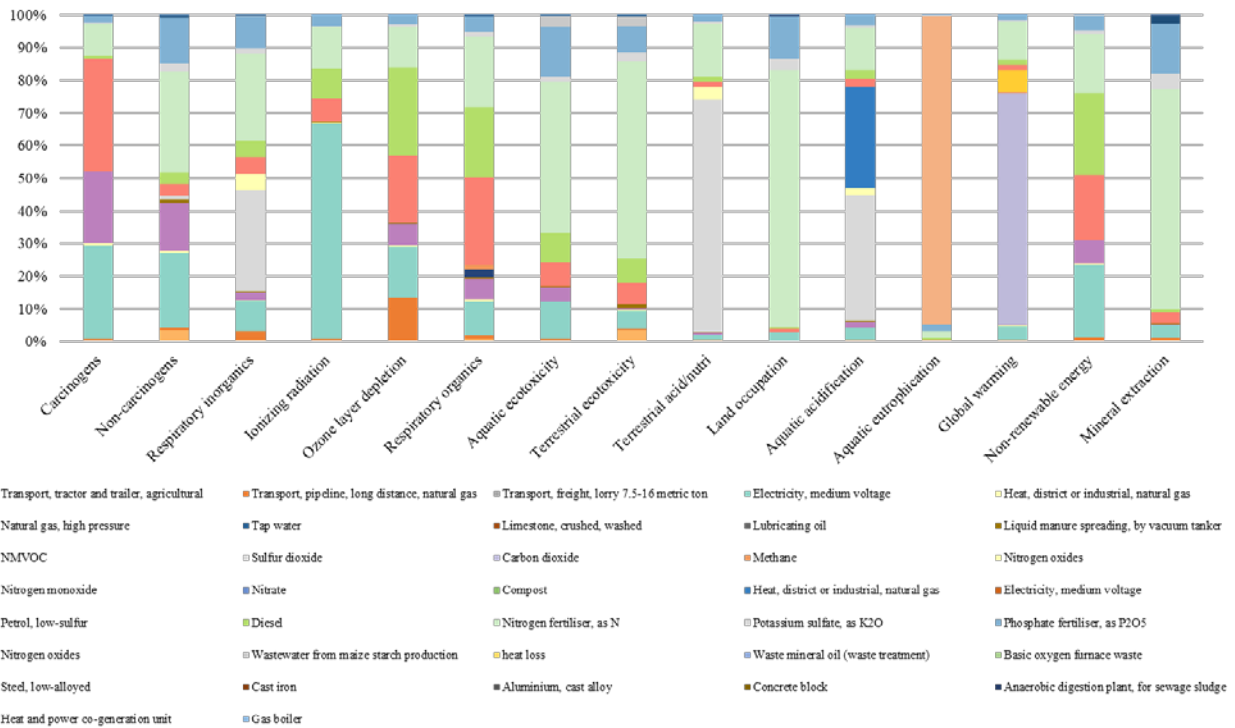
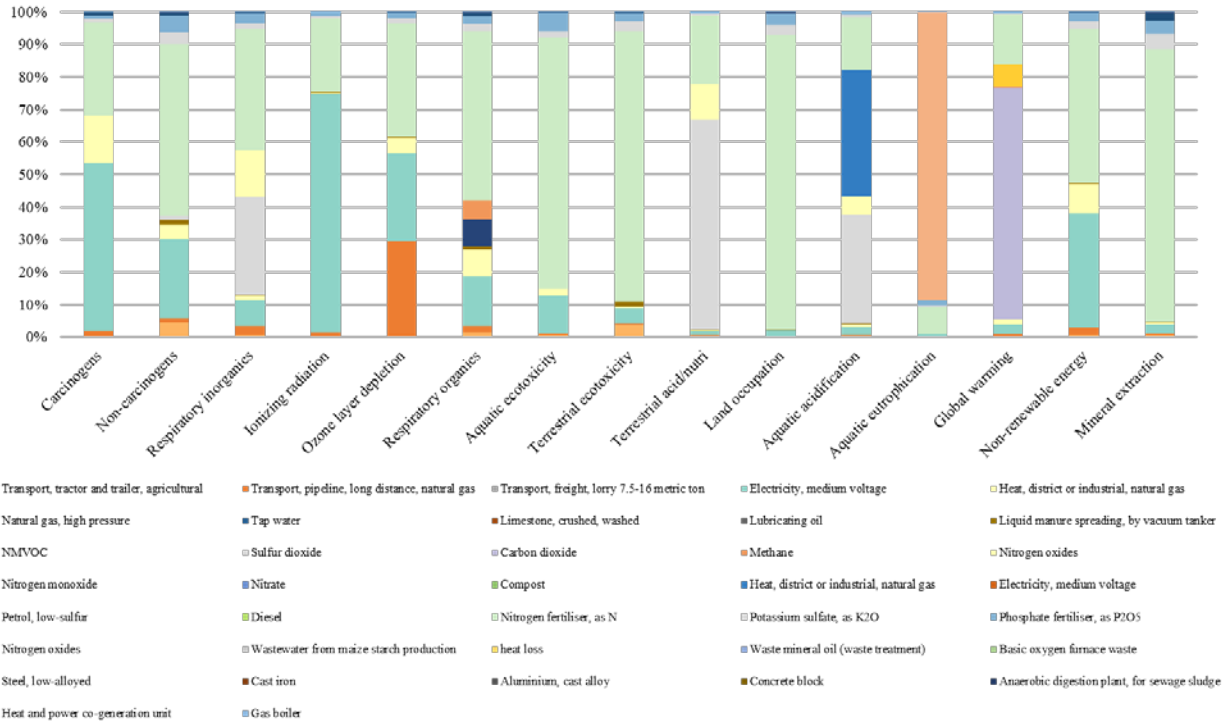
The LCA points results (Figure S3) along with their breakdown in terms of end-point category points is provided in this section. The representation of the end-point category scores in terms of their respective units is also provided (Table S8). Finally, the sensitivity analysis results for the three representative technologies - slow pyrolysis, gasification and HTL are provided (Figure S5).



(a)



(b)



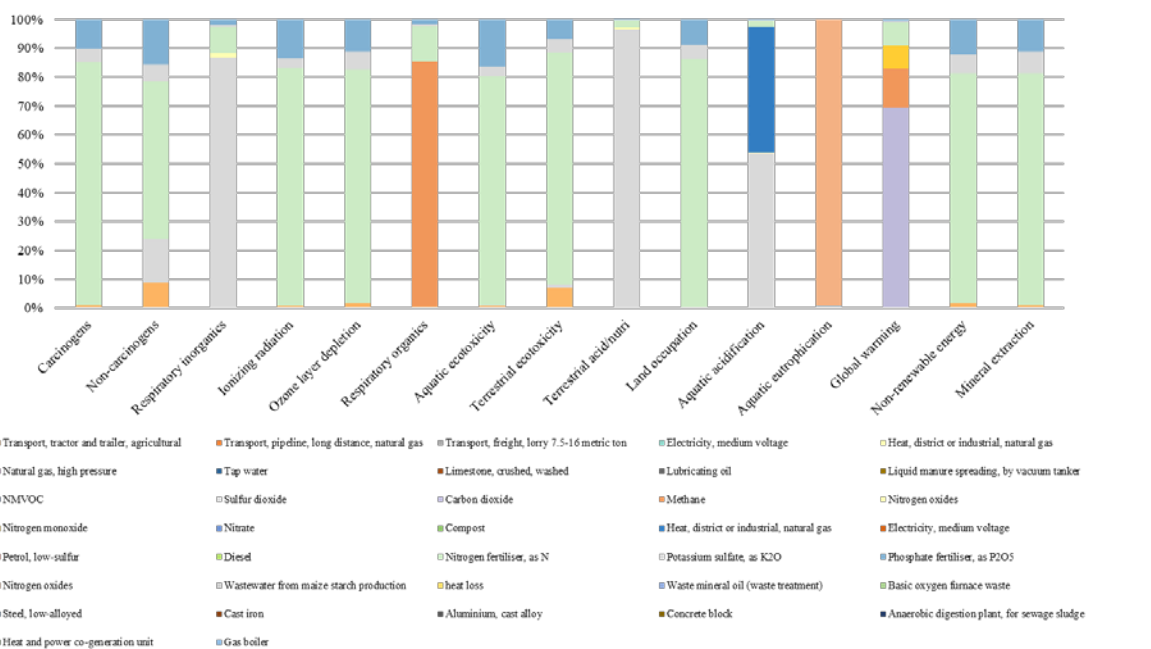
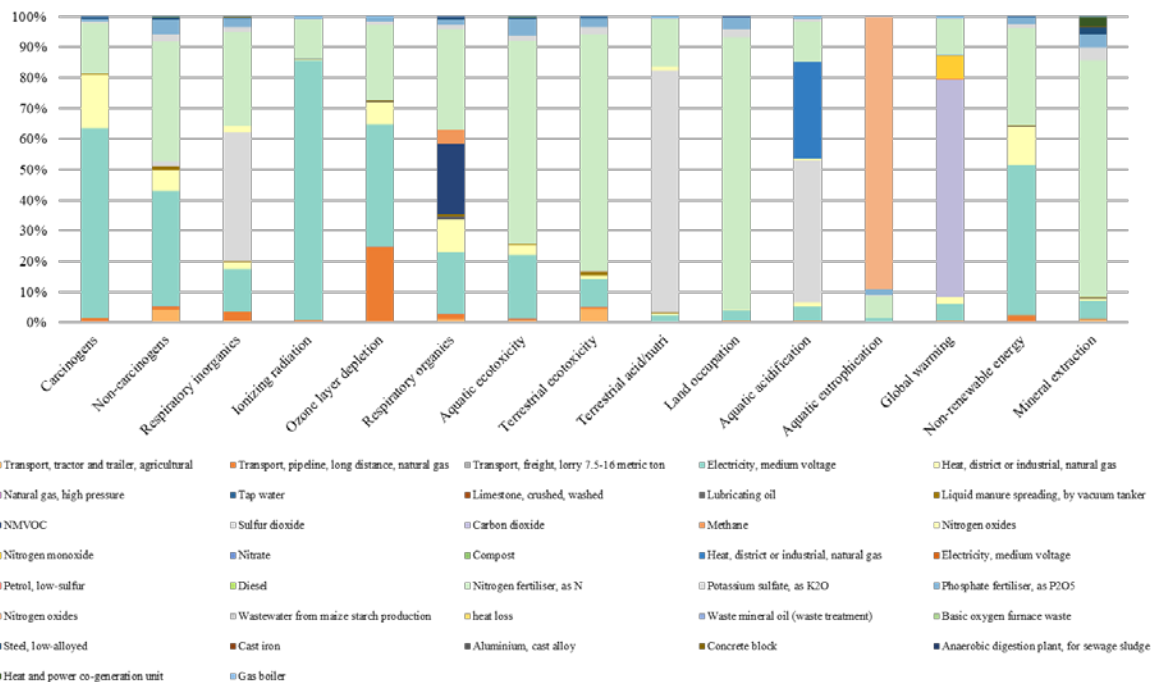


Figure S3. Calculated LCIA midpoint contribution for the different technologies: (a) fast pyrolysis; (b) gasification; (c) HTC; (d) HTL; (e) SCWG; (f) Reference case of land application.

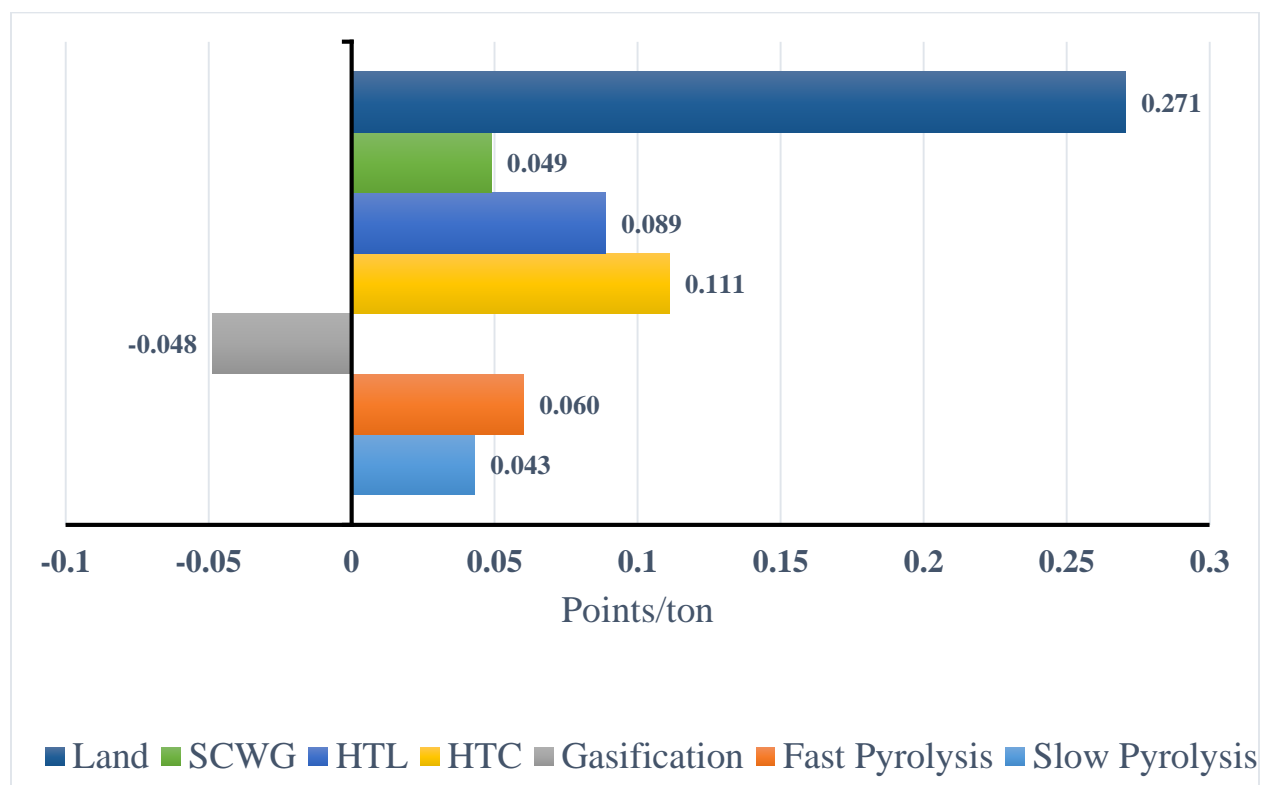


Figure S4. Calculated LCIA normalized points for the different technologies. Lesser value corresponds to better environmental performance.

Table S8. Life cycle impact assessment (LCIA) of thermochemical conversion of poultry litter (values given for wet poultry litter). Lower values correspond to better environmental performance, and we can see that the reference case of land application has the highest values for all five columns, indicating worst performance in each category. The values in the ‘Human Health’ category should be read by multiplying the values in the column by 10^{-3} (this was done to ensure consistency in the number of decimal digits). For example, the human health value for the reference case is 0.000882 DALY/1000 kg wet poultry litter.

	LCIA score (Points/1000 kg)	Human Health ($10^{-3} \cdot$ DALY /1000 kg)	Ecosystem Quality (PDF \cdot m² \cdot year /1000 kg)	Climate Change (Kg CO₂-eq/1000 kg)	Resource Depletion (MJ primary /1000 kg)
Reference case	0.271	0.882	204.6	1410	-1653
Slow pyrolysis	0.043	-0.071	-49.5	658	-1464
Fast pyrolysis	0.060	-0.089	-52.3	978	-3370
Gasification	-0.048	-0.433	-68.8	853	-10,412
HTC	0.111	0.072	-122.7	1170	-1247
HTL	0.089	0.022	-86.7	1192	-4287
SCWG	0.049	-0.095	-100	1079	-5975

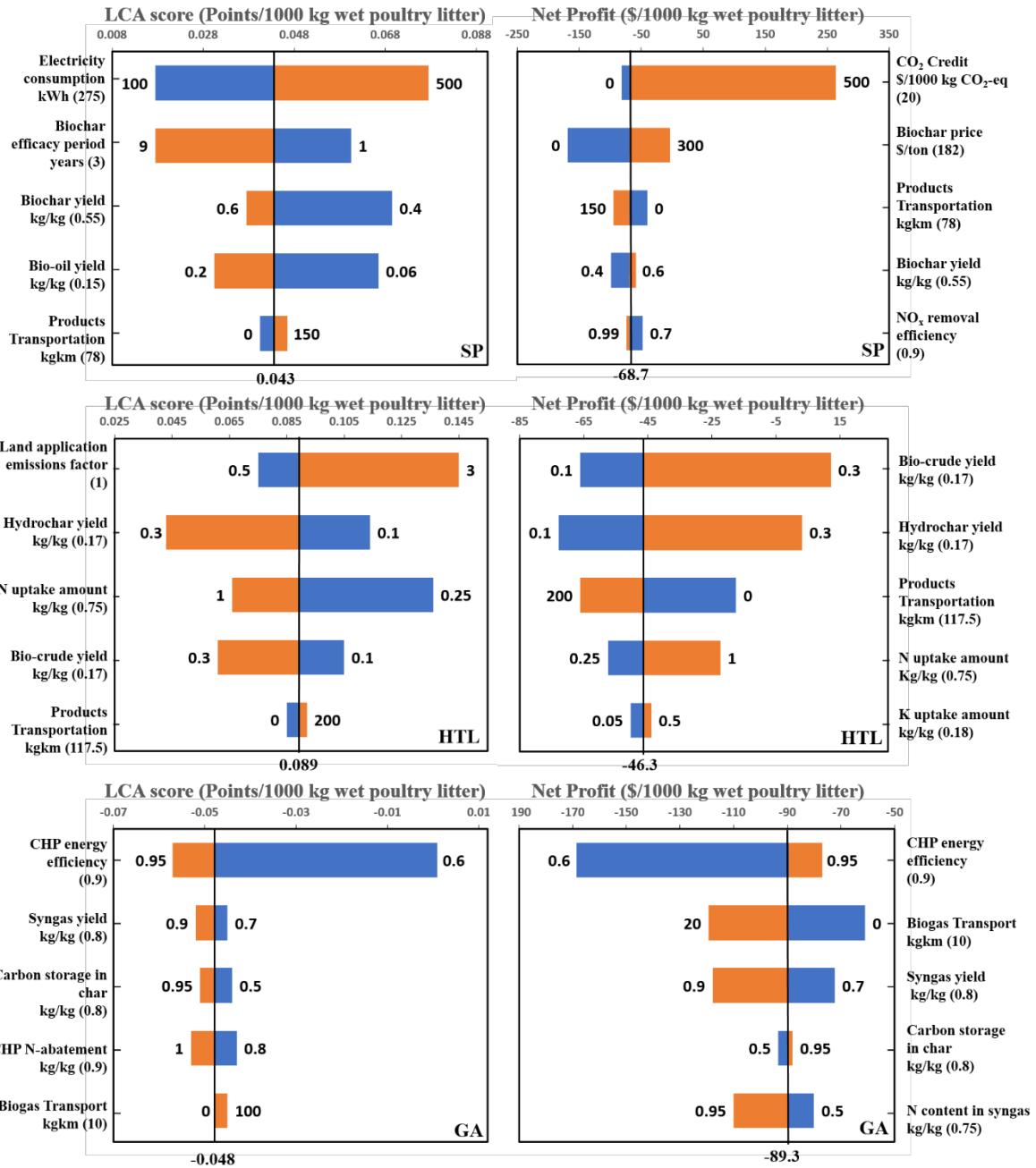


Figure S5. Sensitivity analysis of the LCA (left) and TEA (right) results for the three technologies - slow pyrolysis (SP), hydrothermal liquefaction (HTL) and gasification (GA). The centerline represents the base case results and the numbers in the brackets represent the base values of the input parameters used in our calculations. The numbers next to each bar represent the minimum and maximum values for the change in the input parameter, whereas the numbers on the x-axis

represent the corresponding change in output. The orange and blue bars stand for the increase and decrease of input factors, respectively. For the LCA score, negative points mean better performance. For the TEA results, positive values refer to profits while negative values correspond to losses.

Table S9. Additional sensitivity analysis of the six technologies to highlight the uncertainty in the capital costs and the associated effects on the economic performance. The input variations for the capital costs are different for different technologies as shown in the table.^{50, 51}

Technology	Parameter	Low	Baseline	High	Change (%)
slow pyrolysis	Capital cost (\$/ton)	-35.84	-44.80	-53.76	±20
	Economic performance\$/ton	-45.72	-68.70	-88.38	±31.81
fast pyrolysis	Capital cost (\$/ton)	-61.33	-76.66	-91.99	±20
	Economic performance\$/ton	-94.30	-130.80	-165.76	±27.48
gasification	Capital cost (\$/ton)	-95.11	-118.89	-142.67	±20
	Economic performance\$/ton	-40.52	-89.30	-138.08	±54.62
HTC	Capital cost (\$/ton)	-38.82	-64.70	-90.58	±40
	Economic performance\$/ton	-34.25	-94.62	-154.99	±63.80
HTL	Capital cost (\$/ton)	-22.69	-37.81	-52.93	±40
	Economic performance\$/ton	-10.39	-46.25	-82.11	±77.54
SCWG	Capital cost (\$/ton)	-60.20	-100.34	-140.48	±40
	Economic performance\$/ton	-14.29	-106.61	-198.93	±86.60

References

- (1) Bruun, E. W.; Ambus, P.; Egsgaard, H.; Hauggaard-Nielsen, H., Effects of slow and fast pyrolysis biochar on soil C and N turnover dynamics. *Soil Biol. Biochem.* **2012**, *46*, 73-79, DOI: 10.1016/j.soilbio.2011.11.019
- (2) Kantarli, I. C.; Kabadayi, A.; Ucar, S.; Yanik, J., Conversion of poultry wastes into energy feedstocks. *Waste Manage.* **2016**, *56*, 530-539, DOI: 10.1016/j.wasman.2016.07.019
- (3) Agblevor, F. A.; Beis, S.; Kim, S. S.; Tarrant, R.; Mante, N. O., Biocrude oils from the fast pyrolysis of poultry litter and hardwood. *Waste Manage.* **2010**, *30* (2), 298-307, DOI: 10.1016/j.wasman.2009.09.042
- (4) Kim, S. S.; Agblevor, F. A.; Lim, J., Fast pyrolysis of chicken litter and turkey litter in a fluidized bed reactor. *J. Ind. Eng. Chem.* **2009**, *15* (2), 247-252, DOI: 10.1016/j.jiec.2008.10.004
- (5) Pandey, D. S.; Kwapinska, M.; Gomez-Barea, A.; Horvat, A.; Fryda, L. E.; Rabou, L.; Leahy, J. J.; Kwapinski, W., Poultry Litter Gasification in a Fluidized Bed Reactor: Effects of Gasifying Agent and Limestone Addition. *Energy Fuels* **2016**, *30* (4), 3085-3096, DOI: 10.1021/acs.energyfuels.6b00058
- (6) Woolf, D.; Lehmann, J.; Lee, D. R., Optimal bioenergy power generation for climate change mitigation with or without carbon sequestration. *Nat. Commun.* **2016**, *7*, DOI: 10.1038/ncomms13160
- (7) Huang, Y.; Anderson, M.; McIlveen-Wright, D.; Lyons, G. A.; McRoberts, W. C.; Wang, Y. D.; Roskilly, A. P.; Hewitt, N. J., Biochar and renewable energy generation from poultry litter waste: A technical and economic analysis based on computational simulations. *Appl. Energy* **2015**, *160*, 656-663, DOI: 10.1016/j.apenergy.2015.01.029
- (8) Paturska, A.; Repele, M.; Bazbauers, G., Economic assessment of biomethane supply system based on natural gas infrastructure. In *International Scientific Conference Environmental and Climate Technologies, Conect 2014*, Valtere, S., Ed. 2015; Vol. 72, pp 71-78.
- (9) Jeswani, H. K.; Whiting, A.; Martin, A.; Azapagic, A., Environmental and economic sustainability of poultry litter gasification for electricity and heat generation. *Waste Manage.* **2019**, *95*, 182-191, DOI: 10.1016/j.wasman.2019.05.053
- (10) Font-Palma, C., Characterisation, kinetics and modelling of gasification of poultry manure and litter: An overview. *Energy Convers. Manage.* **2012**, *53* (1), 92-98, DOI: 10.1016/j.enconman.2011.08.017
- (11) Ekpo, U.; Ross, A. B.; Camargo-Valero, M. A.; Williams, P. T., A comparison of product yields and inorganic content in process streams following thermal hydrolysis and hydrothermal processing of microalgae, manure and digestate. *Bioresour. Technol.* **2016**, *200*, 951-960, DOI: 10.1016/j.biortech.2015.11.018
- (12) Mante, O. D.; Agblevor, F. A., Influence of pine wood shavings on the pyrolysis of poultry litter. *Waste Manage.* **2010**, *30* (12), 2537-2547, DOI: 10.1016/j.wasman.2010.07.007
- (13) Shimizu, M.; Hatano, R.; Arita, T.; Kouda, Y.; Mori, A.; Matsuura, S.; Niimi, M.; Jin, T.; Desyatkin, A. R.; Kawamura, O.; Hojito, M.; Miyata, A., The effect of fertilizer and manure application on CH₄ and N₂O emissions from managed grasslands in Japan. *Soil Sci. Plant Nutr.* **2013**, *59* (1), 69-86, DOI: 10.1080/00380768.2012.733926
- (14) Steed, J.; Hashimoto, A. G., Methane emissions from typical manure management-systems. *Bioresour. Technol.* **1994**, *50* (2), 123-130, DOI: 10.1016/0960-8524(94)90064-7

- (15) VanderZaag, A. C.; Jayasundara, S.; Wagner-Riddle, C., Strategies to mitigate nitrous oxide emissions from land applied manure. *Anim. Feed Sci. Technol.* **2011**, *166-67*, 464-479, DOI: 10.1016/j.anifeedsci.2011.04.034
- (16) Smith, P.; Martino, D.; Cai, Z.; Gwary, D.; Janzen, H.; Kumar, P.; McCarl, B.; Ogle, S.; O'Mara, F.; Rice, C.; Scholes, B.; Sirotenko, O.; Howden, M.; McAllister, T.; Pan, G.; Romanenkov, V.; Schneider, U.; Towprayoon, S.; Wattenbach, M.; Smith, J., Greenhouse gas mitigation in agriculture. *Philos. Trans. R. Soc., B* **2008**, *363* (1492), 789-813, DOI: 10.1098/rstb.2007.2184
- (17) Chadwick, D. R.; Pain, B. F.; Brookman, S. K. E., Nitrous oxide and methane emissions following application of animal manures to grassland. *J. Environ. Qual.* **2000**, *29* (1), 277-287, DOI: 10.2134/jeq2000.00472425002900010035x
- (18) Leytem, A. B.; Dungan, R. S.; Bjerneberg, D. L.; Koehn, A. C., Emissions of Ammonia, Methane, Carbon Dioxide, and Nitrous Oxide from Dairy Cattle Housing and Manure Management Systems. *J. Environ. Qual.* **2011**, *40* (5), 1383-1394, DOI: 10.2134/jeq2009.0515
- (19) Hou, Y.; Velthof, G. L.; Oenema, O., Mitigation of ammonia, nitrous oxide and methane emissions from manure management chains: a meta-analysis and integrated assessment. *Glob. Chang. Biol.* **2015**, *21* (3), 1293-1312, DOI: 10.1111/gcb.12767
- (20) Nemecek, T.; Kägi, T.; Blaser, S., Life cycle inventories of agricultural production systems. *Ecoinvent report version* **2007**, *2*, 15,
- (21) Sharpley, A.; Moyer, B., Phosphorus forms in manure and compost and their release during simulated rainfall. *Journal of environmental quality* **2000**, *29* (5), 1462-1469,
- (22) McDowell, R.; Sharpley, A., Variation of phosphorus leached from Pennsylvanian soils amended with manures, composts or inorganic fertilizer. *Agriculture, ecosystems & environment* **2004**, *102* (1), 17-27,
- (23) Angst, T. E.; Sohi, S. P., Establishing release dynamics for plant nutrients from biochar. *GCB Bioenergy* **2013**, *5* (2), 221-226, DOI: 10.1111/gcbb.12023
- (24) Jeffery, S.; Verheijen, F. G. A.; van der Velde, M.; Bastos, A. C., A quantitative review of the effects of biochar application to soils on crop productivity using meta-analysis. *Agric., Ecosyst. Environ.* **2011**, *144* (1), 175-187, DOI: 10.1016/j.agee.2011.08.015
- (25) Ding, Y.; Liu, Y. G.; Liu, S. B.; Li, Z. W.; Tan, X. F.; Huang, X. X.; Zeng, G. M.; Zhou, L.; Zheng, B. H., Biochar to improve soil fertility. A review. *Agron. Sustainable Dev.* **2016**, *36* (2), DOI: 10.1007/s13593-016-0372-z
- (26) Roberts, K. G.; Gloy, B. A.; Joseph, S.; Scott, N. R.; Lehmann, J., Life Cycle Assessment of Biochar Systems: Estimating the Energetic, Economic, and Climate Change Potential. *Environ. Sci. Technol.* **2010**, *44* (2), 827-833, DOI: 10.1021/es902266r
- (27) Kammann, C.; Ratering, S.; Eckhard, C.; Muller, C., Biochar and Hydrochar Effects on Greenhouse Gas (Carbon Dioxide, Nitrous Oxide, and Methane) Fluxes from Soils. *J. Environ. Qual.* **2012**, *41* (4), 1052-1066, DOI: 10.2134/jeq2011.0132
- (28) Bargmann, I.; Rillig, M. C.; Kruse, A.; Greef, J. M.; Kucke, M., Effects of hydrochar application on the dynamics of soluble nitrogen in soils and on plant availability. *J. Plant Nutr. Soil Sci.* **2014**, *177* (1), 48-58, DOI: 10.1002/jpln.201300069
- (29) Kambo, H. S.; Dutta, A., A comparative review of biochar and hydrochar in terms of production, physico-chemical properties and applications. *Renewable Sustainable Energy Rev.* **2015**, *45*, 359-378, DOI: 10.1016/j.rser.2015.01.050

- (30) Tonini, D.; Hamelin, L.; Wenzel, H.; Astrup, T., Bioenergy Production from Perennial Energy Crops: A Consequential LCA of 12 Bioenergy Scenarios including Land Use Changes. *Environ. Sci. Technol.* **2012**, *46* (24), 13521-13530, DOI: 10.1021/es3024435
- (31) Wang, Z.; Chen, D.; Song, X.; Zhao, L., Study on the combined sewage sludge pyrolysis and gasification process: mass and energy balance. *Environ. Technol.* **2012**, *33* (22), 2481-2488,
- (32) Peters, J. F.; Iribarren, D.; Dufour, J., Biomass Pyrolysis for Biochar or Energy Applications? A Life Cycle Assessment. *Environ. Sci. Technol.* **2015**, *49* (8), 5195-5202, DOI: 10.1021/es5060786
- (33) Tews, I. J.; Zhu, Y.; Drennan, C.; Elliott, D. C.; Snowden-Swan, L. J.; Onarheim, K.; Solantausta, Y.; Beckman, D. *Biomass Direct Liquefaction Options. TechnoEconomic and Life Cycle Assessment*; Pacific Northwest National Lab. (PNNL): Richland, WA, United States, 2014,
- (34) Wernet, G.; Bauer, C.; Steubing, B.; Reinhard, J.; Moreno-Ruiz, E.; Weidema, B., The ecoinvent database version 3 (part I): overview and methodology. *Int. J. Life Cycle Assess.* **2016**, *21* (9), 1218-1230, DOI: 10.1007/s11367-016-1087-8
- (35) Yan, B.; Pan, Y.; Yan, J., Spatial Allocation of Animal Manure Nutrient Based on GIS. *J. Indian Soc. Remote Sensing* **2018**, *46* (4), 617-624,
- (36) Wu, H.; Hanna, M. A.; Jones, D. D., Life cycle assessment of greenhouse gas emissions of feedlot manure management practices: Land application versus gasification. *Biomass Bioenergy* **2013**, *54*, 260-266,
- (37) Song, W. P.; Guo, M. X., Quality variations of poultry litter biochar generated at different pyrolysis temperatures. *J. Anal. Appl. Pyrolysis* **2012**, *94*, 138-145, DOI: 10.1016/j.jaap.2011.11.018
- (38) Shabangu, S.; Woolf, D.; Fisher, E. M.; Angenent, L. T.; Lehmann, J., Techno-economic assessment of biomass slow pyrolysis into different biochar and methanol concepts. *Fuel* **2014**, *117*, 742-748, DOI: 10.1016/j.fuel.2013.08.053
- (39) Meyer, S.; Glaser, B.; Quicker, P., Technical, Economical, and Climate-Related Aspects of Biochar Production Technologies: A Literature Review. *Environ. Sci. Technol.* **2011**, *45* (22), 9473-9483, DOI: 10.1021/es201792c
- (40) Jolliet, O.; Margni, M.; Charles, R.; Humbert, S.; Payet, J.; Rebitzer, G.; Rosenbaum, R., IMPACT 2002+: A new life cycle impact assessment methodology. *Int. J. Life Cycle Assess.* **2003**, *8* (6), 324-330, DOI: 10.1007/bf02978505
- (41) Humbert, S.; Schryver, A. D.; Bengoa, X.; Margni, M.; Jolliet, O. *IMPACT 2002+: User Guide (Draft for version Q2.21)*; Quantis: 2012,
- (42) Van Doren, L. G.; Posmanik, R.; Bicalho, F. A.; Tester, J. W.; Sills, D. L., Prospects for energy recovery during hydrothermal and biological processing of waste biomass. *Bioresour. Technol.* **2017**, *225*, 67-74, DOI: 10.1016/j.biortech.2016.11.030
- (43) Daily Prices: Today in Energy - U.S. Energy Information Administration (EIA). <https://www.eia.gov/todayinenergy/prices.php> (accessed 2019-08).
- (44) Weekly On-Highway Diesel Prices - New York State Energy Research and Development Authority (NYSERDA). <https://www.nyserd.ny.gov/Researchers-and-Policymakers/Energy-Prices/On-Highway-Diesel/Weekly-Diesel-Prices> (accessed 2019-08).
- (45) Weekly Average Motor Gasoline Prices - New York State Energy Research and Development Authority (NYSERDA). <https://www.nyserd.ny.gov/Researchers-and-Policymakers/Energy-Prices/Motor-Gasoline/Weekly-Average-Motor-Gasoline-Prices> (accessed 2019-08).

- (46) Berge, N. D.; Flora, J. R.; Drive, B.; Carolina, N., Energy Source Creation from Diverted Food Wastes via Hydrothermal Carbonization. *Raleigh, North Carolina, USA* **2015**,
- (47) Sorrels, J. L., Selective Catalytic Reduction. In *SCR Cost Manual*, US Environmental Protection Agency (EPA): NC, United States, 2016.
- (48) Lantz, M.; Svensson, M.; Björnsson, L.; Börjesson, P., The prospects for an expansion of biogas systems in Sweden—incentives, barriers and potentials. *Energy Policy* **2007**, *35* (3), 1830-1843,
- (49) *State and Trends of Carbon Pricing 2018*; World Bank, Ecofys: Washington, DC, United States, 2018,
- (50) Jones, S. B.; Zhu, Y.; Anderson, D. B.; Hallen, R. T.; Elliott, D. C.; Schmidt, A. J.; Albrecht, K. O.; Hart, T. R.; Butcher, M. G.; Drennan, C.; Snowden-Swan, L. J.; Davis, R.; Kinchin, C. *Process Design and Economics for the Conversion of Algal Biomass to Hydrocarbons: Whole Algae Hydrothermal Liquefaction and Upgrading*; PNNL-23227; Pacific Northwest National Lab. (PNNL): Richland, WA, 2014, DOI: <https://doi.org/10.2172/1126336>
- (51) Wright, M. M.; Satrio, J. A.; Brown, R. C.; Daugaard, D. E.; Hsu, D. D. *Techno-Economic Analysis of Biomass Fast Pyrolysis to Transportation Fuels*; NREL/TP-6A20-46586; National Renewable Energy Lab. (NREL): Golden, CO, 2010, DOI: <https://doi.org/10.2172/993332>